# 2014 CHINOOK SALMON SONAR ENUMERATION ON THE BIG SALMON RIVER 

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#### Abstract

A long range dual frequency identification sonar (DIDSON) was used to enumerate the Chinook salmon escapement to the Big Salmon River in 2014. The sonar was operated on the Big Salmon River for its tenth year at the same site used for the 2005 to 2013 projects; approximately 1.5 km upstream of the confluence of the Yukon River. Sonar operation began on July 11 and continued through to August 18. A total of 6,277 targets identified as Chinook salmon was counted during the period of operation, with 2 fish counted on the first day of operation and 15 fish counted on the last day of operation. Based on linear extrapolation of the last 7 days of sonar counts it is estimated a further 44 Chinook entered the system after sonar operation stopped. This yields a total estimated Chinook escapement of 6,321. The first Chinook salmon passing the Big Salmon sonar station was observed on July 11, the first day of operations. This was 4 days earlier than the earliest fish previously recorded (July 15 in 2005 and 2006). The peak daily count of 421 fish occurred on July 27, at which date $53 \%$ of the estimated run had passed the sonar station (10 days earlier than the 9 year average for $50 \%$ passage). Approximately $90 \%$ of the run had passed the station by August 6 ( 9 days earlier than average). The 2014 Big Salmon count of 6,321 Chinook salmon was $33 \%$ above the previous 9 -year average passage estimate into the system of 4,758 Chinook. Genetic stock identification sampling conducted at the Eagle sonar indicated the Big Salmon River stock group comprised $2.4 \%$ (s.d. $=2.1$ ) of upper Yukon River Chinook salmon escapement in 2014. However, based on the Big Salmon and Eagle sonar counts of 6,321 and 63,431 respectively, the Big Salmon run comprised approximately $10 \%$ of the total above border escapement.

Carcass samples were collected between Aug 15 and Aug 25 over approximately 145 km of the Big Salmon River, yielding 143 Chinook salmon samples. Of the total, 73 ( $51 \%$ ) were female and $70(49 \%)$ were male. The mean MEF length of females and males sampled was 852 mm and 745 mm , respectively. All sampling data and scale cards were submitted to DFO Whitehorse stock assessment; scales were subsequently read by the Pacific Biological Station fish ageing lab. Complete age data was determined from 114 of the Chinook sampled; the remaining 29 samples yielded partial or no ages due to regenerate scales. Age-6 was the dominant age class of females ( $50 \%$ ), followed by age- $5(4 \%)$ and age- $7(1 \%)$. Of the males, age- 5 was the dominant age class $(30 \%)$ followed by ages $6(10 \%), 4(4 \%)$ and $3(2 \%)$.


## INTRODUCTION

The 2014 Big Salmon sonar project marks the tenth year Chinook enumeration has been conducted on this system. The DIDSON sonar units used on the Big Salmon and other escapement enumeration projects have been found to be reliable and provide accurate counts of migrating salmon (Enzhofer et al. 2010, Holmes et al. 2006, Mercer \& Wilson 2006-2014). Due to high flows and wilderness recreation use of the Big Salmon River, the utilization of traditional salmon weir techniques on this river is not feasible. For these reasons the DIDSON sonar was selected as a relatively low impact, non-intrusive method of accurately enumerating annual Chinook escapements to the Big Salmon River system. The use of sonar allows for enumeration of migrating Chinook salmon while minimizing negative impacts on fish behaviour and providing un-restricted recreational use of the river. This report is a summary of the 2014 project.

Based on the 2005 - 2014 sonar operations, the Big Salmon River has been shown to be a significant contributor to upper Yukon River Chinook production. The 2005-2014 average sonar count is 4,914 (range 1,329 to 9,261 ). These counts represented an average of $10.5 \%$ of the total upper Yukon River Chinook spawning escapement point estimate for these years (Unpublished DFO Whitehorse data).

The goal of the program is to provide stock assessment information that will enhance the ability of salmon management agencies to manage Yukon River Chinook salmon. Quantifying Chinook escapement into upper Yukon River index streams allows for independent (from Pilot station and Eagle sonar project estimates) assessment of total above border Chinook escapements. Accurate Chinook escapement enumeration of select tributaries combined with stock composition information could generate upper Yukon River Chinook spawning escapement estimates within quantified statistical parameters.

In addition to the sonar operation, carcass sampling was conducted to obtain age, sex and length data from the 2014 Big Salmon Chinook escapement. This information provides important biological baseline data on the health of the stocks as well as information used in constructing future pre-season run forecasts.

## Study Area

The Big Salmon River flows in a north-westerly direction from the headwaters at Quiet and Big Salmon lakes to its confluence with the Yukon River (Figure 1). The river and its tributaries drain an area of approximately $6,760 \mathrm{~km}^{2}$, predominantly from the Big Salmon Range of the Pelly Mountains. Major tributaries of the Big Salmon River include the North Big Salmon River and the South Big Salmon River. The Big Salmon River can be accessed by boat either from Quiet Lake on the South Canol Road, from the Yukon River on the Robert Campbell and Klondike Highways, or from Lake Laberge via the Thirty Mile and Yukon rivers. The sonar site is at a remote location, approximately 130 air kilometers from Whitehorse. It is accessible by either boat or float plane.

## Objectives

The objectives of the 2014 Big Salmon River sonar project were:

1. To provide an accurate count of the total Chinook salmon escapement in the Big Salmon River using a high resolution DIDSON sonar unit.
2. To conduct a carcass pitch on the Big Salmon River to obtain age-sex-length (ASL) data from as many post-spawned Chinook as possible with a target goal of 5\% of the total run and document egg retention and principal recovery locations.

## METHODS

## Site selection

Sonar operations were set up at the same site used since 2005. This site, located approximately 1.5 km upstream from the confluence with the Yukon River (Figure 1), was initially selected for the following reasons:

- It is a sufficient distance upstream of the mouth to avoid straying or milling Chinook salmon destined for other headwater spawning sites.
- The site is in a relatively straight section of the river and far enough downstream from any bends in the river so that recreational boaters using the river have a clear view of the instream structures.
- The river flow is laminar and swift enough to preclude milling or 'holding' behaviour by migrating fish.
- Bottom substrates consist of gravel and cobble evenly distributed along the width of the river.
- The stream bottom profile allows for complete ensonification of the water column.
- The site is accessible by boat and floatplane.

The physical characteristics of the river at this site have not changed over the 10 years of sonar operation. It is anticipated that this site will continue to be used as long as the sonar program operates.

## Permits

An application was submitted in 2005 to Transport Canada (Marine Branch), Navigable Waters Protection for approval to install partial fish diversion fences in a navigable waterway. Approval was granted for ongoing annual sonar operations as described in the original application.


Figure 1. Big Salmon River Watershed and location of the 2014 Big Salmon sonar station.

## Camp and Sonar Station Set-up

Due to the early run timing in 2014, project mobilization started on July 9. Initial access to the project site and transportation of associated equipment and supplies was by boat from Little Salmon Village. Other supplies and personnel were transported from Whitehorse via floatplane. Subsequent camp access, crew changes, and delivery of supplies and fuel were accomplished either by riverboat or floatplane.

A five year licence of occupation was granted to the contractor in 2009 by the Yukon Territorial Government Lands Branch for the sonar camp (lower Big Salmon River, N $61^{\circ} 52^{\prime} 45^{\prime}$ ', W $134^{\circ}$ $\left.53^{\prime} 08^{\prime \prime}\right)$. This precluded the requirement of annual land use permits and allowed for the construction of upgraded and more permanent facilities at this site. As in 2013 the camp was comprised of two wall tents and one cabin. The cabin was used primarily for storage and the tents were used for accommodation and to house the sonar and computer equipment.

## Diversion Fence Construction

At the onset of the project, fence structures were placed in the river to divert shoreline migrating Chinook salmon into a 36 m migration corridor in the center of the river (Figure 2). Fence construction was initiated on July 10 and completed by July 11. Fence structures were constructed as in previous years using conduit panels and metal tripods stored on site.


Figure 2. Aerial view of sonar station camp and partial weirs, (photo from 2010 project). Blue outline denotes ensonified portion of the river.

## River Profile

A boat mounted Biosonics DTX split beam sonar, aimed $90^{\circ}$ down from the surface, was used to obtain a cross section profile of the river bottom at the sonar site. Data was collected from three bank to bank transects of the river. These transects were located 5 m upstream, at the center and 5 m downstream of the DIDSON sonar beams. The bottom profile was similar for all three transects. The cross section profile where the sonar was deployed is presented in Figure 3. The
cross section profile of the river bottom has remained relatively unchanged since the project started in 2005.


Figure 3. Cross section profile of Big Salmon River at sonar site using a Biosonics DTX split beam echo-sounder.
Note: Top of yellow line is river bottom, thalweg $=1.97 \mathrm{~m}$. Transect view looking down river. The near field of the transducer prevents readings at depths less than 1 m as indicated by the white band.

## Sonar and Software Configuration

The sonar unit was placed next to the south bank at the site used in previous sonar operations (Figures 4). The configuration of the DIDSON sonar support apparatus was similar to that used in previous years at this site. The unit was mounted on an adjustable stand constructed of 2-inch steel galvanized pipe similar in design to those used at other DIDSON sonar projects (Galbreath and Barber 2005). The stand consisted of two T-shaped legs 120 cm in height connected by a 90 cm crossbar. The sonar unit was bolted to a steel plate suspended from the cross bar that was connected to the stand with adjustable fittings (Kee Klamps ${ }^{\text {TM }}$ ). The adjustable clamps allowed the sonar unit to be raised or lowered according to fluctuating water levels as well as enabling rotation of the transducer lens to adjust the beam angle.


Figure 4. Sonar transducer unit and mounting stand in position. (Photo 2011)
The sonar system was powered by two sets of 6 gel cell batteries connected in two parallel circuits to create a 12 volt power source. The battery banks were charged by six 80 watt solar
panels and a backup 2.0 kW generator. An 800 watt inverter was used to obtain 110 volt AC from the batteries to supply power for the computers and the sonar unit. A rotating solar panel platform allowed the panels to be manually rotated to directly face the sun thereby increasing the efficiency of the solar panel array.

A concentrator lens has been attached to the sonar unit during its operation since 2009. This lens reduces the vertical ensonified field from $14^{\circ}$ to $8^{\circ}$, thus increasing the resolution of all target images. The DIDSON sonar produces an ensonified field $29^{\circ}$ wide in the horizontal plane and with the concentrator lens, $8^{\circ}$ deep in the vertical plane. The DIDSON transducer lens was positioned at a depth of approximately 12 cm below the surface of the river and angled downward approximately $3^{\circ}$ from horizontal resulting in the ensonified field of view remaining parallel to the surface of the river.

Using an $8^{\circ}$ lens on a sonar unit deployed horizontally results in a beam depth of 1.05 m at a distance of 7.5 m from the sonar. The average water levels encountered in 2014 allowed for use of the concentrator lens throughout the project. A table was used from simple trigonometry formulae to enable the sonar operators to determine the beam depth for given water depths and sonar window start lengths (Appendix 1).

Once the sonar was in place and positioned, the primary sonar unit settings and software were configured. The receiver gain was set at -40 dB , the window start at 5.86 m , window length at 40 m , and auto frequency enabled for the duration of the project. The recording frame rate was typically set at 4 frames per second, which was the highest frame rate the computers could process with a window length setting of 40 m . Two laptop computers were used for the project, one recording the DIDSON files and one for reviewing the files. All files were saved and placed on a backup 500 GB external hard drive.

## Sonar Data Collection

Sonar recording began on July 11 and continued until August 18. Sonar data was collected continuously and stored automatically in pre-programmed, 20 minute date stamped files. This resulted in an accumulation of 72 files over a 24 hour period. These files were subsequently reviewed the following day and stored on the active PC as well as backed up on the external hard drive.

To optimize target detection during file review, the background subtraction feature was used to remove static images such as the river bottom and weir structures. The intensity (brightness) was set at 40 dB and threshold (sensitivity) at 3dB. The playback speed depended on the preference and experience of the observer, but was generally set between 40 and 50 frames per second, approximately 8 to 10 times the recording rate. When necessary, the recording was stopped when a fish was observed and replayed at a slower rate for positive identification. Chinook salmon images were visually counted using a hand counter and the total count from each file was entered into an excel spreadsheet. A record of each 20 minute file count as well as hourly, daily and cumulative counts was maintained throughout the run.

The target measurement feature of the DIDSON software was used when required to estimate the size of the observed fish. All fish 50 cm and larger were categorized as Chinook. The smallest sampled Big Salmon Chinook during the 2014 carcass pitch was 48.5 cm . Four of the 143 sampled fish ( $2.8 \%$ ) had a MEF length less than 50 cm . The largest target categorized as a resident fish based on size and swimming behaviour was approximately 30 cm .

Fish moving downstream identified as live Chinook were subtracted from each file total. It is assumed Chinook migrating downstream were strays. Straying of migrating salmon is not unusual and temporary ${ }^{1}$ straying has been documented in telemetry studies of Yukon River Chinook (Eiler et al. 2006). The number of assumed strays detected is typically low and in 2014 amounted to 66 fish or $1.1 \%$ of the total run.

## Precision of Fish Counts

It is standard practice in salmon enumeration sonar projects to review a sub-set of recorded data and apply an expansion factor to obtain a total estimate of fish passage. The variance associated with this expansion method can be quantified and incorporated into the total fish passage estimate (Enzenhofer et al., 2010; Crane and Dunbar 2007, 2010). For the Big Salmon sonar project, all recorded files were reviewed in their entirety so there was no variance associated with the expansion of a sub-set of a file data.

The precision of the file counts was measured by double reviewing a sub-set of all the files recorded. Precision refers to the repeatability of a count between different individuals for the same data file. Files for review were randomly selected from each day of sonar operation. Approximately $14 \%$ of the total files were reviewed by a second observer. The re-count from each file was recorded for comparison with the original.

The average percent error (APE) method can be used to quantify the repeatability (precision) of counts, particularly those counts with high fish passage rates (Enzenhofer et. al, 2010). This formula is expressed as:
$A P E=\frac{1}{N} \sum_{j=1}^{N}\left[\frac{1}{R} \sum_{i=1}^{R} \frac{\left|X_{i j}-\bar{X}_{j}\right|}{\bar{X}_{j}}\right] \times 100$
where N is the number of events counted by R observers, Xij is the $i$ th count of the $j$ th event and Xj is the average count of the j th event.

However, because of the relatively low number of fish per hour in most of the Big Salmon sonar files, the percent error could be over-estimated. For example, if the first counter observed 2 upstream fish and the second counter missed one, the APE would be as high as $33 \%$. This is because of the leverage that small numerical differences in low counts have on the overall calculation of APE. For this reason, the average percent error for this project was calculated using files with fish counts $\geq 5$ fish/ file.

As well as calculating APE, a sample variance estimator based on the absolute difference between readers was used to quantify the precision of the counts and the net variability between readers.

[^0]
## Cross Section Distribution

The position of each Chinook observed within the cross section profile of the river was recorded in 5 m increments. This provided a range frequency histogram illustrating the cross sectional pattern of migrating Chinook.

## Carcass Pitch

The upper reaches of the Big Salmon River were accessed using a 6.0 m open skiff powered by a 60 hp outboard jet motor. An initial carcass pitch trip was started on August 19 but had to be halted due to motor problems. After the motor was repaired, the crew made one extended trip upriver on August 22 through August 25. Carcass pitch efforts extended from the camp approximately 145 river kilometers to the first logjam located 20 km downstream from Big Salmon Lake. In addition to collection of ASL data, information was collected on the egg retention of the sampled females. The principal locations of the recovered carcasses and moribund fish were also recorded.

The carcass pitch involved collecting dead and moribund Chinook using a spear and sampling each fish. Carcass sampling consisted of collecting five scales per fish and placing them in prescribed scale cards. The sex and mid-eye-fork and post-orbital hypural lengths (to the nearest 0.5 cm ) were also recorded for each recovered fish.

## RESULTS

## Chinook Salmon Counts

The first Chinook salmon was observed on July 11, on the first day of operations. The peak daily count of 421 fish occurred on July 27, at which time $53 \%$ of the estimated run had passed the sonar station ( 10 days earlier than the 9 year average for $50 \%$ passage); $90 \%$ of the run had passed the station by August 6 ( 9 days earlier than average). Daily and cumulative counts are presented in Appendix 2 and Figure 5. A total of 6,277 targets identified as Chinook salmon was counted past the sonar station from July 11 through to August 18. Because the sonar was removed before the run was totally complete, the counts were estimated for an additional 6 days after sonar operations were stopped. This was done through extrapolation of the previous 6 days of the sonar counts based on the linear regression $y=-2.5429 x+31.4$. This extrapolation added 44 fish to bring the total count to 6,321 .

The 2014 daily counts exhibited a normal distribution. The run timing was approximately 8 days earlier than the average run timing observed in the previous 9 years (Figure 5). Daily counts from 2005 through 2014 are in listed in Appendix 3.

## Precision of Fish Counts

Of the 2,788 sonar files recorded and analysed, a total of 394 (14\%) were reviewed by a second observer (Table 1). Of the 394 files reviewed, a total of 14 files ( $3.5 \%$ ) exhibited a discrepancy between readers. Of the 14 files that exhibited inconsistencies between readers, an additional 16 fish were observed by the reviewer and 6 fish missed by the reviewer. This would yield a net gain of 10 fish for the 394 files that were reviewed. Expansion of this subset of files to cover the
total number of files recorded would result in a possible total of 71 Chinook ( $1.1 \%$ of the total run) that may not have been observed and counted.


Figure 5. Daily counts of Chinook salmon passing the Big Salmon River sonar station in 2014 and 2005 through 2013.

Figure 6 illustrates the relationship between counts of 2 different file readers using daily pooled original and reviewed files. Linear regression between readers showed variation between counts but overall the correlation is high $\left(\mathrm{R}^{2}=0.99\right)$.


Figure 6. Linear regression between daily pooled sonar file counts that had been analysed by 2 different readers.
Note: Data points are daily pooled initial file counts (y axis) and reviewed file counts (x axis).
The average percent error (APE) was calculated for the 151 reviewed files that had fish counts $\geq$ 5 fish/file. The APE for this subset was $0.17 \%$.

Table 1. Double reviewed files and calculated difference between counts.

|  | Count | \% |
| :--- | :--- | :--- |
| Total files recorded during project | 2,788 |  |
| Total files double reviewed | 394 | $14 \%$ |
| Total fish counted in double reviewed files | 1007 |  |
| Total files with + variance | 10 | $3.1 \%$ |
| Total files with - variance | 4 | $1.2 \%$ |
| Total Files with variance | 14 | $4.3 \%$ |
| Total plus fish | 16 | $+1.6 \%$ |
| Total minus fish | -6 | $-0.6 \%$ |
| Difference | +10 | $+1.0 \%$ |

## Cross Section Distribution

The cross sectional distribution pattern of the migrating Chinook as detected by the sonar is presented in Figure 7. The largest proportion of fish migrated near the south bank in deeper water at a distance of 15-20 meters from the sonar. It should be noted the distribution likely does not reflect the typical in-river migration pattern as the weir structures channel the fish into the 36 m wide opening.

The water levels experienced in 2014 were considered average which may account for the typical migration pattern. The water levels and temperatures recorded during the project are listed in Appendix 4.


Figure 7. 2010-2014 Big Salmon River Chinook range/frequency in cross section profile.
Note: The $0-7 \mathrm{~m}$ range from the sonar has a deflection fence in place.

## Carcass Pitch

A total of 143 dead or moribund Chinook was recovered during the carcass pitch. Of the fish sampled, $73(51 \%)$ were female and $70(49 \%)$ were male. The mean fork length of females and males sampled was 852 mm and 745 mm , respectively. The length frequency of Chinook
sampled is presented in Figure 8. Complete age data ${ }^{2}$ was determined from 114 of the Chinook sampled; the remaining 29 samples yielded partial or no ages due to regenerate scales. Age-6 $(1.4)^{3}$ was the dominant age class of females ( $50 \%$ ), followed by age- 5 (1.3) (4\%) and age-7 (1.2) ( $1 \%$ ). Of the males, age 5 (1.3) was the dominant age class ( $30 \%$ ) followed by ages 6 (1.4) (10\%), 4 (1.2) (4\%) and 3(1.1) (2\%). Mean length and age data with complete age information is presented in Table 2.


Figure 8. Length/frequency histogram of Big Salmon Chinook sampled in 2014.

Table 2. Age, length, and sex of Chinook sampled from the Big Salmon River, 2014.

| SEX | AGE | Mean MEF | Count | \% |
| :--- | :--- | :--- | :--- | :--- |
| Female | 1.3 | 819 | 5 | $4 \%$ |
|  | 1.4 | 852 | 57 | $50 \%$ |
|  | 1.5 | 825 | 1 | $1 \%$ |
| Female total |  | 848 | 63 | $55 \%$ |
| Male | 1.1 | 533 | 2 | $2 \%$ |
|  | 1.2 | 610 | 4 | $4 \%$ |
|  | 1.3 | 737 | 34 | $30 \%$ |
|  | 1.4 | 837 | 11 | $10 \%$ |
| Male total |  | 741 | 51 | $45 \%$ |
| Total |  | 798 | 114 | $100 \%$ |

Egg retention of sampled dead and moribund female Chinook was low. Of the 73 females sampled, two ( $2.7 \%$ ) were not fully spawned out. The egg retention in these two fish was estimated to be approximately $50 \%$. Complete age, length and sex data as well as egg retention and principal recovery locations are presented in Appendix 5.

[^1]
## DISCUSSION

The 2014 Big Salmon sonar project was successful in enumerating the Chinook salmon passing the station and there is a high level of confidence the sonar count accurately reflected the Chinook escapement into the system. Water levels were moderate throughout most of the season and the migration corridor of 36 m was maintained throughout the sonar operation. The DIDSON sonar unit and related power and computer equipment functioned well throughout the project. Short interruptions in sonar recording due to maintenance or power interruptions resulted in a total of 5 hours and 28 minutes recording loss over the course of the project.

The comparison of the counts of files reviewed by two different individuals exhibited a high degree of precision between both reviewers ${ }^{4}$. Repeated counts of the files were observed to produce the same counts $96.5 \%$ of the time for all files read. Average percent error of all the reviewed files was low $(0.17 \%)$. Since most of the discrepancies involved missed fish it can be surmised that sonar counts would be biased low. The variability between readers was not factored into the daily counts and the resultant potentially missed fish (71) were not added to the total sonar count.

The 2014 Eagle sonar project on the Yukon River downstream of the Canada/U.S. border yielded a spawning escapement ${ }^{5}$ estimate of 63,331 Chinook salmon (DFO Whitehorse unpublished data 2015). Based on both the Big Salmon and Eagle Chinook sonar counts, the Big Salmon stock contributed $10.0 \%$ of the total above border Chinook production in 2014.

Genetic stock identification (GSI) samples were also obtained at the Eagle sonar site using drift nets. The GSI data provides information on the stock composition of the total above border Chinook escapement. The 2014 mean un-weighted proportional contribution of the Big Salmon River stock to the Chinook border escapement based on analysis of the GSI samples was $2.4 \%$, (SD 2.1\%) (DFO Whitehorse unpublished data). The 2014 proportional contribution of the Big Salmon River stock to the Chinook above border escapement based on analysis of the GSI samples was significantly lower than was observed from 2005 through 2013 (mean 9.0\%; range $6.4 \%-14.0 \%$ ). In addition, 3 years of radio telemetry studies (2002-2004) indicated the Big Salmon contribution to the Chinook above border escapement averaged $11.4 \%$ (range $9.2-15.1$; Appendix 6). Moreover, as noted above, the 2014 GSI derived stock proportion is significantly lower than the proportion derived from the Big Salmon and Eagle Chinook sonar counts (10.0\%).

Using Big Salmon sonar counts and the proportion of Big Salmon origin stock derived from the GSI sampling, an expanded Chinook border escapement estimate could be calculated. However, due to the relatively large standard deviation associated with the GSI derived stock proportion, the precision of the estimate would be poor and an expanded escapement estimate based on the Big Salmon count and GSI data would have limited value. It is possible to obtain a Chinook border escapement with better precision based on analysis of the GSI samples and sonar counts from the Teslin sonar project. These results are presented in the 2014 Teslin River sonar project report (Mercer 2015).

To date, neither the Big Salmon/Eagle sonar counts nor the GSI based stock proportions serve as useful predictive indices on the run strength of the stocks. Appendix 7 illustrates the relationship between the Eagle sonar counts and the Big Salmon sonar counts from 2005 through 2014. As

[^2]expected there is a relationship between the Big Salmon sonar counts and JTC estimates (mark/recapture and Eagle sonar), however, the relationship is non-significant $\left(R^{2}=0.61\right)$. It can be expected that there will be inter- annual variance in the proportional contribution of the Big Salmon stock group to the total above border escapement as evidenced by the GSI and telemetry derived stock proportions (Appendix 6).

Appendix 8 illustrates the GSI based expansion of Big Salmon sonar counts and the JTC above border escapement estimates from 2005 to 2014. There is no significant correlation between the Big Salmon Chinook sonar counts and the calculated Big Salmon escapement estimates derived from the Eagle sonar count and GSI proportions $\left(\mathrm{R}^{2}=0.28\right)$. The low 2014 GSI derived estimate is an obvious outlier and is due to the relatively low 2014 GSI stock proportion (2.4\%). Even with this outlier removed, the correlation has limited predictive value and does not yield sufficient precision to be used as an escapement index or to monitor long term trends for the system. Hence, neither the Big Salmon nor the Eagle sonar counts could be considered redundant in this context. More years of data and the elimination of suspected outlier years may increase the predictive precision of this method. It is worth noting that based on the Eagle and Big Salmon sonar counts, the Big Salmon stock GSI proportions appear to be biased low (Appendix 9). This may be a result of non-representative genetic sampling at Eagle and/or error in the genetic identification of the Big Salmon stock.
The number of samples collected in the 2014 carcass pitch was lower than expected. This was due in part to motor problems during the first carcass pitch trip. Repairs to the motor took four days, after which there was time for only one carcass pitch trip.

It is recognized that potential biases can occur with carcass sampling, just as biases can and do occur with all salmon sampling methods currently employed other than those that sample the whole population. Chinook dead pitch sampling bias has been examined in several studies (Mears and Dubois 2009; Zou 2002). Within a given year, carcass pitch sampling biases can be influenced by environmental factors (turbidity, flows), timing of data collection in relation to the die off period, inter-annual variation in age and sex structure of the population, sample size, sampling location and the experience of the samplers. As a generalization, age and sex compositions from survey samples underestimated the ages of small fish and males while overestimating those of large fish and females. Carcass pitch sampling is bias toward females because of their overall larger size and their propensity to live longer and stay on redds until they expire. Mears study indicated males were underrepresented in the carcass survey by $14 \%$ and females were overrepresented by $20 \%$ (Mears and Dubois 2009). Zou (2002) found in a multiyear study that males were under represented by $8 \%$ and females over represented by $12 \%$. However, indications are that the population age class structures typically are representative of the population. In summary, the carcass survey data may be able to provide an adequate estimate of age class structure for each sex, but the numbers of individuals per sex may deviate from the whole population. It is worth noting those age classes that contribute the highest productive potential (age 5 and age 6) have the lowest probability of exhibiting age class sampling bias.

The Big Salmon program has been ongoing for ten consecutive years. There is value in maintaining an upper Yukon Chinook escapement monitoring project that provides accurate data over a long time series. The rationale for continuing this project is:

- It has proven to be a viable and consistent means of obtaining accurate escapement counts as well as age, sex and length data of Chinook salmon returning to the Big Salmon River.
- The Big Salmon stock comprises, on average, approximately 10 to $11 \%$ of the total upper Yukon Chinook escapement; the fourth highest stock composition behind the Yukon Mainstem, the Pelly and the Teslin systems.
- There is now one full generation of escapement data for the Big Salmon stock. Continuation of the project will provide ensuing recruitment information on those escapements. The development of biologically based escapement goals is typically based on stock recruitment modelling. These models are based on escapement estimates incorporating a long time series. The importance of long time series and continuous data sets related to escapement monitoring cannot be over emphasized. The data from this project has been an investment for the YRP and management agencies to date. To halt the project now will diminish the value of the investment the YRP and management agencies have put into the project to date.
- Big Salmon escapement information coupled with GSI stock composition data can provide an independent annual estimate of the total above border Chinook spawning escapement as well as provide information on the precision of the GSI stock proportions.


## ACKNOWLEDGEMENTS

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Appendix 1. Sonar beam depth at tilt angles $0^{\circ}-45^{\circ}$ and with start window lengths 6.7 m and 7.5 m .

| 8 Degree Lens |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Horizontal Beam |  |  | 6.67 m start window |  |  | 7.5 m start window |  |  |
| $\begin{aligned} & \text { Distance From } \\ & \text { Sonar (m) } \end{aligned}$ | Depth of Beam (m) | Width of Beam (m) | Tilt Degree @ 6.67 m from sonar | $\begin{aligned} & \text { Depth Added } \\ & (\mathrm{m}) \end{aligned}$ | TOTAL DEPTH <br> @ 6.67m (m) | Tilt Degree @ 7.5m from sonar | Depth Added (m) | $\begin{gathered} \text { TOTAL DEPTH @ } \\ 7.5 \mathrm{~m}(\mathrm{~m}) \end{gathered}$ |
| 2.00 | 0.28 | 1.03 | 1.00 | 0.06 | 0.99 | 1.00 | 0.07 | 1.00 |
| 3.00 | 0.42 | 1.55 | 2.00 | 0.12 | 1.05 | 2.00 | 0.14 | 1.07 |
| 4.00 | 0.56 | 2.07 | 3.00 | 0.18 | 1.11 | 3.00 | 0.20 | 1.13 |
| 5.00 | 0.70 | 2.59 | 4.00 | 0.24 | 1.17 | 4.00 | 0.27 | 1.20 |
| 6.00 | 0.84 | 3.10 | 5.00 | 0.30 | 1.23 | 5.00 | 0.34 | 1.27 |
| 6.67 | 0.93 | 3.45 | 6.00 | 0.36 | 1.29 | 6.00 | 0.41 | 1.34 |
| 7.00 | 0.98 | 3.62 | 7.00 | 0.42 | 1.35 | 7.00 | 0.48 | 1.41 |
| 7.50 | 1.05 | 3.88 | 8.00 | 0.48 | 1.41 | 8.00 | 0.55 | 1.48 |
| 8.00 | 1.12 | 4.14 | 9.00 | 0.55 | 1.48 | 9.00 | 0.61 | 1.54 |
| 9.00 | 1.26 | 4.66 | 10.00 | 0.61 | 1.54 | 10.00 | 0.68 | 1.61 |
| 10.00 | 1.40 | 5.17 | 11.00 | 0.67 | 1.60 | 11.00 | 0.75 | 1.68 |
| 11.00 | 1.54 | 5.69 | 12.00 | 0.73 | 1.66 | 12.00 | 0.82 | 1.75 |
| 12.00 | 1.68 | 6.21 | 13.00 | 0.80 | 1.73 | 13.00 | 0.90 | 1.83 |
| 13.00 | 1.82 | 6.72 | 14.00 | 0.86 | 1.79 | 14.00 | 0.97 | 1.90 |
| 14.00 | 1.96 | 7.24 | 15.00 | 0.92 | 1.85 | 15.00 | 1.04 | 1.97 |
| 15.00 | 2.10 | 7.76 | 16.00 | 0.99 | 1.92 | 16.00 | 1.11 | 2.04 |
| 16.00 | 2.24 | 8.28 | 17.00 | 1.05 | 1.98 | 17.00 | 1.19 | 2.12 |
| 17.00 | 2.38 | 8.79 | 18.00 | 1.12 | 2.05 | 18.00 | 1.26 | 2.19 |
| 18.00 | 2.52 | 9.31 | 19.00 | 1.19 | 2.12 | 19.00 | 1.34 | 2.27 |
| 19.00 | 2.66 | 9.83 | 20.00 | 1.26 | 2.19 | 20.00 | 1.41 | 2.34 |
| 20.00 | 2.80 | 10.34 | 21.00 | 1.32 | 2.25 | 21.00 | 1.49 | 2.42 |
| 21.00 | 2.94 | 10.86 | 22.00 | 1.39 | 2.32 | 22.00 | 1.57 | 2.50 |
| 22.00 | 3.08 | 11.38 | 23.00 | 1.46 | 2.39 | 23.00 | 1.65 | 2.58 |
| 23.00 | 3.22 | 11.90 | 24.00 | 1.54 | 2.47 | 24.00 | 1.73 | 2.66 |
| 24.00 | 3.36 | 12.41 | 25.00 | 1.61 | 2.54 | 25.00 | 1.81 | 2.74 |
| 25.00 | 3.50 | 12.93 | 26.00 | 1.68 | 2.61 | 26.00 | 1.89 | 2.82 |
| 26.00 | 3.64 | 13.45 | 27.00 | 1.76 | 2.69 | 27.00 | 1.98 | 2.91 |
| 27.00 | 3.78 | 13.97 | 28.00 | 1.83 | 2.76 | 28.00 | 2.06 | 2.99 |
| 28.00 | 3.92 | 14.48 | 29.00 | 1.91 | 2.84 | 29.00 | 2.15 | 3.08 |
| 29.00 | 4.06 | 15.00 | 30.00 | 1.99 | 2.92 | 30.00 | 2.24 | 3.17 |
| 30.00 | 4.20 | 15.52 | 31.00 | 2.07 | 3.00 | 31.00 | 2.33 | 3.26 |
| 31.00 | 4.34 | 16.03 | 32.00 | 2.16 | 3.09 | 32.00 | 2.42 | 3.35 |
| 32.00 | 4.48 | 16.55 | 33.00 | 2.24 | 3.17 | 33.00 | 2.52 | 3.45 |
| 33.00 | 4.62 | 17.07 | 34.00 | 2.33 | 3.26 | 34.00 | 2.62 | 3.55 |
| 34.00 | 4.76 | 17.59 | 35.00 | 2.42 | 3.35 | 35.00 | 2.72 | 3.65 |
| 35.00 | 4.89 | 18.10 | 36.00 | 2.51 | 3.44 | 36.00 | 2.82 | 3.75 |
| 36.00 | 5.03 | 18.62 | 37.00 | 2.60 | 3.53 | 37.00 | 2.92 | 3.85 |
| 37.00 | 5.17 | 19.14 | 38.00 | 2.70 | 3.63 | 38.00 | 3.03 | 3.96 |
| 38.00 | 5.31 | 19.65 | 39.00 | 2.79 | 3.72 | 39.00 | 3.14 | 4.07 |
| 39.00 | 5.45 | 20.17 | 40.00 | 2.89 | 3.82 | 40.00 | 3.26 | 4.19 |
| 40.00 | 5.59 | 20.69 | 41.00 | 3.00 | 3.93 | 41.00 | 3.37 | 4.30 |
| 41.00 | 5.73 | 21.21 | 42.00 | 3.11 | 4.04 | 42.00 | 3.49 | 4.42 |
| 42.00 | 5.87 | 21.72 | 43.00 | 3.22 | 4.15 | 43.00 | 3.62 | 4.55 |
| 43.00 | 6.01 | 22.24 | 44.00 | 3.33 | 4.26 | 44.00 | 3.75 | 4.68 |
| 44.00 | 6.15 | 22.76 | 45.00 | 3.45 | 4.38 | 45.00 | 3.88 | 4.81 |
| 45.00 | 6.29 | 23.28 |  |  |  |  |  |  |

Appendix 2. 2014 daily and cumulative counts of Chinook salmon at the Big Salmon River sonar site.

| DATE | DAILY <br> COUNT | CUMULATIVE | COMMENTS |
| :--- | :---: | :---: | :---: |
| Jul-11 | 2 | 2 | sonar recording starts at 15:00 |
| Jul-12 | 18 | 20 |  |
| Jul-13 | 52 | 72 |  |
| Jul-14 | 52 | 124 |  |
| Jul-15 | 64 | 188 |  |
| Jul-16 | 90 | 278 |  |
| Jul-17 | 115 | 393 |  |
| Jul-18 | 170 | 563 |  |
| Jul-19 | 199 | 762 |  |
| Jul-20 | 236 | 998 |  |
| Jul-21 | 229 | 1227 |  |
| Jul-22 | 284 | 1511 |  |
| Jul-23 | 345 | 1856 |  |
| Jul-24 | 343 | 2199 |  |
| Jul-25 | 356 | 2555 |  |
| Jul-26 | 372 | 2927 |  |
| Jul-27 | 421 | 3348 | peak daily count |
| Jul-28 | 307 | 3655 |  |
| Jul-29 | 380 | 4035 |  |
| Jul-30 | 330 | 4365 |  |
| Jul-31 | 256 | 4621 |  |
| Aug-1 | 207 | 4828 |  |
| Aug-2 | 207 | 5035 |  |
| Aug-3 | 192 | 5227 |  |
| Aug-4 | 190 | 5417 |  |
| Aug-5 | 170 | 5587 |  |
| Aug-6 | 120 | 5707 |  |
| Aug-7 | 114 | 5821 |  |
| Aug-8 | 96 | 5917 |  |
| Aug-9 | 68 | 5985 |  |
| Aug-10 | 61 | 6046 |  |
| Aug-11 | 50 | 6096 |  |
| Aug-12 | 46 | 6142 |  |
| Aug-13 | 25 | 6167 |  |
| Aug-14 | 30 | 6197 |  |
| Aug-15 | 24 | 6221 |  |
| Aug-16 | 24 | 6245 |  |
| Aug-17 | 17 | 6262 |  |
| Aug-18 | 15 | 6277 | sonar recording ends at 24:00 |
| Aug-24 |  | 6321 | Extrapolated count |
|  |  |  |  |
|  |  | 2 |  |

Appendix 3. Daily and average Chinook counts in the Big Salmon River, 2005-2014.

| DATE | Daily <br> Count <br> 2005 | Daily <br> Count <br> 2006 | Daily <br> Count <br> 2007 | Daily <br> Count <br> 2008 | Daily <br> Count <br> 2009 | $\begin{aligned} & \hline \text { Daily } \\ & \text { Count } \\ & 2010 \end{aligned}$ | $\begin{aligned} & \hline \text { Daily } \\ & \text { Count } \\ & \hline 2011 \end{aligned}$ | $\begin{aligned} & \hline \text { Daily } \\ & \text { Count } \\ & \hline 2012 \end{aligned}$ | $\begin{gathered} \hline \text { Daily } \\ \text { Count } \\ \hline 2013 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Daily } \\ \text { Count } \\ \hline 2014 \end{gathered}$ | Daily Average |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11-Jul |  |  |  |  |  |  |  |  |  | 2 | 2 |
| 12-Jul |  |  |  |  |  |  |  |  |  | 18 | 18 |
| 13-Jul | 0 |  |  |  |  |  |  |  |  | 52 | 26 |
| 14-Jul | 0 |  |  |  |  |  |  |  |  | 52 | 26 |
| 15-Jul | 2 | 1 |  |  |  |  |  |  |  | 64 | 22 |
| 16-Jul | 12 | 0 | 2 | 0 |  |  |  |  | 0 | 90 | 17 |
| 17-Jul | 13 | 1 | 0 | 0 |  |  | 2 |  | 0 | 115 | 19 |
| 18-Jul | 23 | 0 | 2 | 0 | 0 |  | 7 | 0 | 0 | 170 | 22 |
| 19-Jul | 13 | 0 | 5 | 1 | 11 |  | 13 | 0 | 0 | 199 | 27 |
| 20-Jul | 23 | 1 | 5 | 0 | 22 | 0 | 15 | 0 | 0 | 236 | 30 |
| 21-Jul | 36 | 3 | 7 | 0 | 47 | 7 | 24 | 0 | 1 | 229 | 35 |
| 22-Jul | 58 | 8 | 11 | 0 | 68 | 14 | 24 | 0 | 1 | 284 | 47 |
| 23-Jul | 92 | 11 | 18 | 1 | 85 | 12 | 43 | 0 | 2 | 345 | 61 |
| 24-Jul | 130 | 21 | 26 | 2 | 135 | 7 | 44 | 0 | 4 | 343 | 71 |
| 25-Jul | 158 | 20 | 52 | 1 | 201 | 12 | 50 | 1 | 3 | 356 | 85 |
| 26-Jul | 204 | 53 | 88 | 3 | 226 | 14 | 56 | 1 | 11 | 372 | 103 |
| 27-Jul | 219 | 95 | 153 | 5 | 346 | 27 | 105 | 1 | 25 | 421 | 140 |
| 28-Jul | 287 | 146 | 237 | 9 | 498 | 46 | 160 | 3 | 44 | 307 | 174 |
| $29-\mathrm{Jul}$ | 290 | 230 | 287 | 9 | 532 | 83 | 192 | 15 | 86 | 380 | 210 |
| 30-Jul | 299 | 321 | 337 | 29 | 594 | 123 | 218 | 12 | 83 | 330 | 235 |
| 31-Jul | 279 | 368 | 400 | 21 | 808 | 141 | 218 | 23 | 150 | 256 | 266 |
| 01-Aug | 333 | 357 | 435 | 23 | 578 | 159 | 260 | 62 | 196 | 207 | 261 |
| 02-Aug | 346 | 379 | 331 | 18 | 715 | 182 | 313 | 76 | 220 | 207 | 279 |
| 03-Aug | 303 | 358 | 304 | 16 | 725 | 216 | 417 | 138 | 264 | 192 | 293 |
| 04-Aug | 292 | 413 | 258 | 31 | 595 | 226 | 426 | 156 | 262 | 190 | 285 |
| 05-Aug | 331 | 496 | 210 | 51 | 559 | 215 | 396 | 196 | 261 | 170 | 289 |
| 06-Aug | 214 | 490 | 178 | 55 | 452 | 221 | 400 | 228 | 225 | 120 | 258 |
| 07-Aug | 188 | 464 | 147 | 78 | 364 | 227 | 317 | 192 | 191 | 114 | 228 |
| 08-Aug | 232 | 464 | 59 | 61 | 295 | 242 | 294 | 235 | 195 | 96 | 217 |
| 09-Aug | 234 | 360 | 74 | 70 | 270 | 248 | 243 | 183 | 156 | 68 | 191 |
| 10-Aug | 203 | 349 | 90 | 98 | 209 | 183 | 160 | 154 | 132 | 61 | 164 |
| 11-Aug | 124 | 348 | 82 | 122 | 183 | 207 | 170 | 106 | 134 | 50 | 153 |
| 12-Aug | 126 | 324 | 98 | 107 | 146 | 174 | 143 | 130 | 113 | 46 | 141 |
| 13-Aug | 125 | 243 | 77 | 109 | 118 | 181 | 100 | 110 | 101 | 25 | 119 |
| 14-Aug | 72 | 196 | 74 | 89 | 117 | 134 | 85 | 81 | 77 | 30 | 96 |
| 15-Aug | 57 | 180 | 66 | 78 | 65 | 114 | 89 | 80 | 65 | 24 | 82 |
| 16-Aug | 40 | 172 | 56 | 70 | 55 | 82 | 63 | 94 | 57 | 24 | 71 |
| 17-Aug | 53 | 104 | 40 | 49 | 63 | 80 | 35 | 70 | 34 | 17 | 55 |
| 18-Aug | 47 | 69 | 64 | 45 | 55 | 53 | 20 | 50 | 32 | 15 | 45 |
| 19-Aug | 35 | 87 | 37 | 17 | 43 | 40 | 18 | 44 | 21 | 14 | 36 |
| 20-Aug | 29 | 59 | 47 | 18 | 35 | 24 | 21 | 38 | 28 | 1 | 31 |
| 21-Aug | 26 | 45 | 11 | 15 | 28 | 18 | 11 | 27 | 20 | \% | 21 |
| 22-Aug | 19 | 50 | 16 | 16 | 14 | 38 | 2 | 19 | 10 | 6 | 19 |
| 23-Aug | 17 | 12 | 23 | 9 | 4 | 24 | 2 | 19 | 14 | 3 | 13 |
| 24-Aug | 13 | 10 | 17 | 2 |  | 20 |  | 14 | 11 | - | 11 |
| 25-Aug | \% |  | 14 | 1 |  | 17 |  | 9 | 6 |  | 9 |
| 26-Aug | 6 |  | 14 |  |  | 6 |  | 6. | 4 |  | 7 |
| 27-Aug | 4, |  | +13 |  |  |  |  | O 0 - | + |  | 6 |
| 28-Aug | < $\mathbf{O}$ - |  | - 1 L |  |  |  |  | \$ $\mathbf{3}$, | - |  | 4 |
| 29-Aug | - ${ }^{\text {a }}$ |  | - |  |  |  |  | 2- |  |  | 6 |
| 30-Aug | - |  | - 8 |  |  |  |  | ¢ |  |  | 5 |
| 31-Aug | - ${ }^{\text {+ }}$ |  | \% |  |  |  |  |  |  |  | 6 |
| 01-Sep |  |  | 4 |  |  |  |  |  |  |  | 4 |
| 02-Sep |  |  | 3. |  |  |  |  |  |  |  | 3 |
| TOTAL: | 5618 | 7308 | 4506 | 1329 | 9261 | 3817 | 5156 | 2584 | 3242 | 6321 |  |

Note: Stippled values were obtained through extrapolation of counts. Shaded areas denote start and end of sonar recording

## Appendix 4. 2014 Big Salmon River water and weather conditions.

| DATE | TIME | AIR TEMP. | $\begin{gathered} \hline \text { WATER } \\ \text { TEMP. }\left({ }^{\circ} \mathbf{C}\right) \end{gathered}$ | WATER LEVEL (cm) | COMMENTS |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 11-Jul | 9:00 AM | - | - | - | Mostly sunny |
| 12-Jul | 7:30 AM | 11.0 | 12.0 | 60 | Cloudy with sunny breaks and light showers |
| 13-Jul | 8:30 AM | 11.0 | 13.0 | 58 | Sunny with cloudy periods |
| 14-Jul | 8:30 AM | 11.0 | 13.0 | 55 | Mostly sunny |
| 15-Jul | 8:00 AM | 10.0 | 13.5 | 52 | Mostly sunny in morning with clouds developing in the afternoon |
| 16-Jul | 8:20 AM | 9.0 | 13.5 | 49 | Mostly cloudy |
| 17-Jul | 8:30 AM | 10.0 | 13.0 | 45 | Mostly sunny with thunder showers in the afternoon |
| 18-Jul | 7:40 AM | 7.0 | 12.5 | 42 | Mostly sunny with showers in the afternoon clearing in the evening |
| 19-Jul | 7:55 AM | 6.0 | 13.0 | 41 | Mostly sunny with showers in the afternoon. |
| 20-Jul | 8:05 AM | 8.0 | 13.0 | 39 | Mostly cloudy with light showers in the afternoon |
| 21-Jul | 8:30 AM | 7.0 | 12.5 | 36 | Cloudy in the morning clearing in the afternoon |
| 22-Jul | 7:50 AM | 5.0 | 12.0 | 34 | Sunny all day |
| 23-Jul | 7:40 AM | 8.0 | 13.0 | 33 | Sunny all day |
| 24-Jul | 7:00 AM | 5.0 | 12.0 | 30 | Cloudy with sunny breaks clearing in the evening |
| 25-Jul | 7:45 AM | 7.0 | 13.0 | 27 | Mostly cloudy with sunny breaks clearing in the evening |
| 26-Jul | 8:05 AM | 6.0 | 11.5 | 25 | Mostly cloudy |
| 27-Jul | - | - | - | 23 | Mostly cloudy with thunderstorm - clearing in the evening |
| 28-Jul | 7:30 AM | 5.0 | 11.5 | 22 | Mix Sun and cloud, showers in evening, heavy rain overnight |
| 29-Jul | 7:50 AM | 8.0 | 12.0 | 22 | Cloudy with showers, rain all night |
| 30-Jul | 7:50 AM | 8.0 | 12.0 | 25 | Rain all day clearing in the evening |
| 31-Jul | 8:00 AM | 7.0 | 11.0 | 65 | Clearing in morning, sunny day. River rises to 85 cm by afternoon |
| 01-Aug | 8:10 AM | 2.0 | 10.0 | 75 | Sunny all day River drops to 65 by evening |
| 02-Aug | 8:10 AM | 7.0 | 11.0 | 56 | Sunny and hot. |
| 03-Aug | 7:50 AM | 5.0 | 11.5 | 45 | Mix of sun and cloud |
| 04-Aug | 8:00 AM | 11.0 | 13.0 | 40 | Mix of sun and cloud |
| 05-Aug | 8:10 AM | 6.0 | 11.5 | 34 | Mix of sun and cloud |
| 06-Aug | 8:00 AM | 6.0 | 12.0 | 31 | Mostly cloudy, small showers in the evening |
| 07-Aug | 7:50 AM | 7.0 | 11.5 | 28 | Mostly cloudy, afternoon showers, heavy wind/squalls, rain in evening |
| 08-Aug | 7:30 AM | 7.0 | 11.0 | 26.5 | Morning fog, mixed sun and cloud during day |
| 09-Aug | 7:30 AM | 7.5 | 11.0 | 24 | Overcast all day. |
| 10-Aug | 7:30 AM | 8.0 | 11.0 | 21.5 | partly cloudy in AM, Overcast in day |
| 11-Aug | 7:30 AM | 12.0 | 12.5 | 20 | morning cloud, then mostly sunny all day |
| 12-Aug | 7:30 AM | 10.5 | 12.5 | 19 | morning showers, clearing in afternoon |
| 13-Aug | 7:50 AM | 6.5 | 12.0 | 18 | Mostly cloudy with clearing in the evening |
| 14-Aug | 8:00 AM | 8.0 | 12.0 | 18 | Mostly cloudy with showers in the evening |
| 15-Aug | 7:45 AM | 11.0 | 11.5 | 16 | Mostly cloudy with a shower, clearing in the evening |
| 16-Aug | 8:30 AM | 7.0 | 11.5 | 14 | Sunny in morning then mostly cloudy with small showers. |
| 17-Aug | 8:05 AM | 9.0 | 10.5 | 12 | Cloudy with showers clearing somewhat in evening, |
| 18-Aug | 7:50 AM | 8.0 | 10.0 | 13 | showers and rain all day |
| 19-Aug | 8:30 AM | - | - | - | Rain in the Morning, leave for dead pitch |

Appendix 5 (a). Age, sex, and length of sampled Chinook on the Big Salmon River, 2014.

| DATE | FISH \# | SEX | \% Spawned (Females) | MEF (mm) | POHL (mm) | AGE* | Recovery Site |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 15-Aug | 1 | M |  | 710 | 625 | 1.3 | 13 |
| 19-Aug | 2 | F | N/A | 785 | 685 | 1.4 | 1 |
| 19-Aug | 3 | M |  | 740 | 655 | 1 F | 1 |
| 19-Aug | 4 | M |  | 755 | 675 | RG | 1 |
| 19-Aug | 5 | F | N/A | 885 | 790 | 1.3 | 1 |
| 19-Aug | 6 | M |  | 620 | 540 | 1.3 | 1 |
| 19-Aug | 7 | M |  | 805 | 735 | 1.3 | 1 |
| 19-Aug | 8 | F | N/A | 820 | 725 | 1.4 | 1 |
| 22-Aug | 9 | M |  | 945 | 820 | 1.4 | 13 |
| 23-Aug | 10 | M |  | 580 | 515 | 1.1 | 13 |
| 23-Aug | 11 | M |  | 920 | 805 | 1.4 | 13 |
| 23-Aug | 12 | M |  | 765 | 690 | 1.3 | 1 |
| 23-Aug | 13 | M |  | 720 | 625 | 1.3 | 1 |
| 23-Aug | 14 | M |  | 530 | 455 | 1.2 | 1 |
| 23-Aug | 15 | M |  | 710 | 620 | 1.3 | 1 |
| 23-Aug | 16 | M |  | 635 | 550 | 1.3 | 1 |
| 23-Aug | 17 | M |  | 950 | 845 | RG | 1 |
| 23-Aug | 18 | F | 100 | 885 | 785 | 1.4 | 1 |
| 23-Aug | 19 | M |  | 735 | 640 | 1F | 2 |
| 23-Aug | 20 | M |  | 920 | 805 | 1.3 | 2 |
| 23-Aug | 21 | M |  | 835 | 730 | 1.3 | 2 |
| 23-Aug | 22 | M |  | 820 | 710 | M4 | 2 |
| 23-Aug | 23 | F | 100 | 850 | 760 | 1.4 | 2 |
| 23-Aug | 24 | M |  | 690 | 605 | 1.3 | 2 |
| 23-Aug | 25 | M |  | 825 | 730 | 1.3 | 2 |
| 23-Aug | 26 | F | 100 | 795 | 705 | 1.4 | 2 |
| 23-Aug | 27 | F | 100 | 845 | 760 | 1.4 | 2 |
| 23-Aug | 28 | M |  | 685 | 605 | 1.3 | 2 |
| 23-Aug | 29 | M |  | 645 | 575 | 1.3 | 2 |
| 23-Aug | 30 | F | 100 | 845 | 760 | 1.4 | 2 |
| 23-Aug | 31 | M |  | 680 | 590 | 1.3 | 2 |
| 23-Aug | 32 | M |  | 900 | 790 | 1.3 | 3 |
| 23-Aug | 33 | F | 100 | 870 | 770 | 1F | 3 |
| 23-Aug | 34 | F | 100 | 725 | 640 | 1.3 | 3 |
| 23-Aug | 35 | F | 100 | 930 | 810 | RG | 3 |
| 23-Aug | 36 | M |  | 725 | 630 | 1F | 3 |
| 23-Aug | 37 | F | 100 | 930 | 820 | 1.4 | 3 |
| 23-Aug | 38 | F | 100 | 870 | 765 | 1.4 | 3 |
| 23-Aug | 39 | M |  | 750 | 660 | 1.3 | 3 |
| 23-Aug | 40 | M |  | 785 | 700 | RG | 3 |
| 23-Aug | 41 | M |  | 665 | 580 | 1.4 | 3 |
| 23-Aug | 42 | M |  | 790 | 695 | 1.4 | 3 |
| 23-Aug | 43 | M |  | 855 | 770 | M3 | 3 |
| 23-Aug | 44 | M |  | 730 | 650 | 1.3 | 3 |
| 23-Aug | 45 | M |  | 645 | 570 | 1.3 | 3 |
| 23-Aug | 46 | M |  | 795 | 705 | 1F | 4 |
| 23-Aug | 47 | M |  | 940 | 820 | 1.4 | 4 |
| 23-Aug | 48 | M |  | 685 | 600 | 1.3 | 4 |
| 23-Aug | 49 | M |  | 700 | 605 | 1.3 | 4 |
| 23-Aug | 50 | M |  | 775 | 685 | 1.3 | 4 |


| DATE | FISH \# | SEX | \% Spawned (Females) | MEF (mm) | POHL (mm) | AGE* | Recovery Site |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 23-Aug | 51 | M |  | 705 | 620 | 1.3 | 4 |
| 23-Aug | 52 | M |  | 785 | 695 | 1.4 | 4 |
| 23-Aug | 53 | M |  | 630 | 555 | 1F | 4 |
| 23-Aug | 54 | F | 100 | 870 | 770 | 1.4 | 4 |
| 23-Aug | 55 | F | 100 | 830 | 745 | 1.4 | 4 |
| 23-Aug | 56 | M |  | 550 | 475 | M2 | 4 |
| 23-Aug | 57 | M |  | 655 | 570 | 1.3 | 4 |
| 23-Aug | 58 | M |  | 705 | 615 | 1.3 | 4 |
| 24-Aug | 59 | M |  | 715 | 620 | 1.4 | 5 |
| 24-Aug | 60 | F | 100 | 815 | 725 | 1.3 | 5 |
| 24-Aug | 61 | F | 100 | 850 | 750 | 1.4 | 5 |
| 24-Aug | 62 | M |  | 485 | 425 | 1.1 | 5 |
| 24-Aug | 63 | M |  | 735 | 640 | 1.2 | 5 |
| 24-Aug | 64 | M |  | 865 | 760 | 1.3 | 5 |
| 24-Aug | 65 | F | 100 | 855 | 755 | 1.4 | 5 |
| 24-Aug | 66 | M |  | 865 | 760 | 1.4 | 5 |
| 24-Aug | 67 | M |  | 665 | 580 | 1.3 | 5 |
| 24-Aug | 68 | M |  | 635 | 555 | 1.2 | 5 |
| 24-Aug | 69 | M |  | 960 | 855 | 1.3 | 6 |
| 24-Aug | 70 | M |  | 1010 | 890 | M4 | 6 |
| 24-Aug | 71 | F | N/A | 810 | 725 | 1.4 | 6 |
| 24-Aug | 72 | F | 100 | N/M | 735 | 1.4 | 6 |
| 24-Aug | 73 | M |  | 815 | 730 | 1.3 | 6 |
| 24-Aug | 74 | F | 100 | 895 | 800 | 1.4 | 6 |
| 24-Aug | 75 | M |  | 945 | 830 | M4 | 6 |
| 24-Aug | 76 | M |  | 805 | 700 | 1.3 | 6 |
| 24-Aug | 77 | F | 100 | 875 | 775 | 1.4 | 6 |
| 24-Aug | 78 | M |  | 875 | 760 | 1.4 | 7 |
| 24-Aug | 79 | M |  | 740 | 645 | 1.4 | 7 |
| 24-Aug | 80 | M |  | 970 | 845 | 1.4 | 7 |
| 24-Aug | 81 | F | 100 | 855 | 770 | 1.4 | 7 |
| 24-Aug | 82 | F | 100 | 910 | 810 | 1.4 | 7 |
| 24-Aug | 83 | F | 100 | 835 | 750 | 1.4 | 7 |
| 24-Aug | 84 | F | 100 | 785 | 690 | 1.4 | 7 |
| 24-Aug | 85 | F | 100 | 900 | 790 | M4 | 7 |
| 24-Aug | 86 | F | 100 | 830 | 745 | 1.4 | 7 |
| 24-Aug | 87 | F | 100 | 870 | 780 | 1.4 | 7 |
| 24-Aug | 88 | F | 100 | 850 | 755 | M4 | 7 |
| 24-Aug | 89 | F | 100 | 820 | 720 | 1.4 | 7 |
| 24-Aug | 90 | F | 50 | 880 | 790 | 1.4 | 7 |
| 24-Aug | 91 | M |  | 540 | 470 | 1.2 | 7 |
| 24-Aug | 92 | F | 100 | 870 | 770 | M3 | 7 |
| 24-Aug | 93 | F | 100 | 925 | 825 | 1.4 | 7 |
| 24-Aug | 94 | F | 100 | 905 | 795 | M3 | 7 |
| 24-Aug | 95 | F | 50 | 920 | 810 | 1.4 | 7 |
| 24-Aug | 96 | M |  | 590 | 520 | M2 | 7 |
| 24-Aug | 97 | F | 100 | 955 | 850 | 1.4 | 7 |
| 24-Aug | 98 | F | 100 | 825 | 750 | 1.3 | 8 |
| 24-Aug | 99 | F | 100 | N/M | 745 | 1.4 | 8 |
| 24-Aug | 100 | F | 100 | 845 | 755 | 1.4 | 8 |


| DATE | FISH \# | SEX | \% Spawned (Females) | MEF (mm) | POHL (mm) | AGE* | Recovery Site |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 24-Aug | 101 | F | 100 | 835 | 740 | 1.4 | 8 |
| 24-Aug | 102 | F | 100 | N/M | 745 | 1.4 | 8 |
| 24-Aug | 103 | F | 100 | N/M | 830 | 1.4 | 8 |
| 24-Aug | 104 | F | 100 | N/M | 725 | 1.4 | 8 |
| 24-Aug | 105 | F | 100 | 905 | 790 | 1.4 | 8 |
| 24-Aug | 106 | F | 100 | 830 | 740 | 1.4 | 8 |
| 24-Aug | 107 | F | 100 | 900 | 800 | 1.4 | 8 |
| 24-Aug | 108 | F | 100 | 880 | 785 | 1.4 | 8 |
| 24-Aug | 109 | F | 100 | 790 | 695 | 1.4 | 8 |
| 24-Aug | 110 | F | 100 | 835 | 755 | 1.4 | 8 |
| 24-Aug | 111 | F | 100 | 840 | 740 | RG | 8 |
| 24-Aug | 112 | F | 100 | N/M | 670 | M4 | 8 |
| 24-Aug | 113 | F | 100 | 810 | 735 | 1.4 | 9 |
| 24-Aug | 114 | F | 100 | 925 | 835 | M4 | 9 |
| 24-Aug | 115 | F | 100 | 845 | 760 | 1.4 | 9 |
| 24-Aug | 116 | M |  | 685 | 610 | M2 | 9 |
| 24-Aug | 117 | F | 100 | 805 | 715 | 1.4 | 9 |
| 24-Aug | 118 | M |  | 735 | 650 | RG | 9 |
| 24-Aug | 119 | M |  | 670 | 580 | M3 | 9 |
| 24-Aug | 120 | F | 100 | N/M | 725 | 1.4 | 9 |
| 24-Aug | 121 | F | 100 | 825 | 740 | 1.5 | 9 |
| 24-Aug | 122 | M |  | 720 | 640 | M3 | 9 |
| 24-Aug | 123 | F | 100 | 840 | 750 | 1.4 | 9 |
| 24-Aug | 124 | F | 100 | 910 | 820 | 1.4 | 9 |
| 24-Aug | 125 | M |  | 705 | 610 | 1.3 | 9 |
| 24-Aug | 126 | M |  | 765 | 670 | 1.3 | 9 |
| 24-Aug | 127 | M |  | 670 | 585 | 1.3 | 9 |
| 24-Aug | 128 | F | 100 | 830 | 755 | 1.4 | 9 |
| 25-Aug | 129 | F | 100 | 845 | 740 | 1.3 | 10 |
| 25-Aug | 130 | F | 100 | 885 | 780 | 1.4 | 10 |
| 25-Aug | 131 | M |  | 700 | 625 | M3 | 10 |
| 25-Aug | 132 | F | 50 | 825 | 725 | 1.4 | 10 |
| 25-Aug | 133 | F | 100 | 880 | 775 | 1.4 | 10 |
| 25-Aug | 134 | F | 100 | 785 | 690 | 1.4 | 10 |
| 25-Aug | 135 | M |  | 730 | 635 | 1.3 | 10 |
| 25-Aug | 136 | F | 100 | 860 | 760 | 1.4 | 10 |
| 25-Aug | 137 | F | 100 | 855 | 760 | 1.4 | 10 |
| 25-Aug | 138 | F | 100 | 900 | 800 | 1.4 | 10 |
| 25-Aug | 139 | F | 100 | 870 | 770 | 1.4 | 10 |
| 25-Aug | 140 | F | 100 | 785 | 710 | 1.4 | 13 |
| 25-Aug | 141 | F | 100 | 805 | 730 | 1 F | 13 |
| 25-Aug | 142 | F | 100 | 730 | 645 | 1.4 | 13 |
| 25-Aug | 143 | M |  | 590 | 500 | 1.3 | 13 |

*European age format; e.g. 1.3 denotes a 5 year old fish with $1+$ years freshwater residence and 3 years marine residence
$\mathrm{RG}=$ Regenerate scale (center is missing from scale)
$\mathrm{M}=$ Marine stage
F = Freshwater stage
N/A = Partially decomposed or consumed, no assessment.
$\mathrm{N} / \mathrm{M}=$ No measurement - fish partially consumed or decomposed

Appendix 5 (b). Primary locations of sampled carcasses and moribund fish recovered on the Big Salmon River, 2014.

| Recovery <br> Site | ${ }^{*}$ GPS Coordinates |
| :---: | :--- |
| 1 | N 61 45.392' |
|  | W 134 37.701' |
|  |  |
| 2 | N 61 40.735' |
|  | W 134 30.664' |
|  |  |
| 3 | N 61 38.066' |
|  | W 134 29.019' |
| 4 | N 61 35.380' |
|  | W 134 21.218' |
|  |  |
| 5 | N 61 32.739' |
|  | W 134 12.157' |
|  | N 61 31.973' |
| 6 | W 134 02.918' |
| 7 | N 61 31.668' |
|  | W 133 58.003' |
|  |  |
| 8 | N 61 31.951' |
|  | W 133 52.588' |
|  | N 61 37.071' |
| 9 | W 133 45.705' |
|  |  |
| 10 | N 61 34.650' |
|  | W 133 38.882' |
|  |  |
| 13 | N 61 52.461 |
|  | W 134 53.129 |
|  |  |

Appendix 6. Estimated proportion of Big Salmon River Chinook and Yukon River Chinook border escapement, 2002 through 2014.

| Year | Method | Estimated \% proportion of border escapement based on telemetry or GSI sampling | Big Salmon sonar count | Border escapement based on Eagle sonar count or mark/recapture | Border escapement based on Big Salmon sonar count and GSI stock proportion |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | Telemetry | 9.2 | n/a | n/a | n/a |
| 2003 | Telemetry | 15.1 | n/a | n/a | n/a |
| 2004 | Telemetry | 10.0 | n/a | n/a | n/a |
| 2005 | Fishwheel GSI Sampling | 10.8 | 5,618 | $67,985^{\text {c }}$ | 52,019 |
| 2006 | Fishwheel GSI Sampling | 9.7 | 7,308 | $62,630^{\text {c }}$ | 75,340 |
| 2007 | Fishwheel GSI Sampling | 10.6 | 4,506 | 34,904 ${ }^{\text {b }}$ | 42,509 |
| 2008 | Fishwheel GSI Sampling | 9.3 | 1,431 | $33,883{ }^{\text {b }}$ | 15,387 |
| 2009 | Gillnet GSI Sampling | 16.9 | 9,261 | $65,278{ }^{\text {b }}$ | 54,799 |
| 2010 | Gillnet GSI Sampling | 11.7 | 3,817 | $32,010^{\text {b }}$ | 32,624 |
| 2011 | Gillnet GSI Sampling | 9.2 | 5,156 | 50,780 ${ }^{\text {a }}$ | 56,043 |
| 2012 | Gillnet GSI Sampling | 6.7 | 2,594 | 32,658 ${ }^{\text {a }}$ | 38,104 |
| 2013 | Gillnet GSI Sampling | 6.6 | 3,239 | 28,669 | 49,136 |
| 2014 | Gillnet GSI Sampling | 2.4 | 6,321 | 63,331 | 263,375 |
| Mean |  | 11.3 | 4,770 | 45,422 | 61,757 |
| Std. Dev. |  | 3 | 2,278 | 15,259 | 65,681 |

${ }^{\text {a }}$ Eagle sonar above border spawning escapement estimate (DFO Whitehorse, unpublished data).
${ }^{\mathrm{b}}$ Eagle sonar estimate (JTC 2012 and Unpublished DFO Whitehorse data).
${ }^{\text {c }}$ Mark/recapture estimate (JTC 2012).
Sources: Osborne et al. 2003; Mercer and Eiler 2004; Mercer 2005; JTC reports 2005 through 2012; unpublished DFO Whitehorse data.

Appendix 7. Big Salmon sonar counts and the JTC above border escapement estimates, 2005 2014.


Appendix 8. GSI based expansion of Big Salmon sonar counts and the JTC above border escapement estimates, 2005 - 2014.


Appendix 9. GSI based estimates of Big Salmon Chinook escapements and comparisons with actual Big Salmon sonar counts, 2005-2014.

| Year | GSI Stock <br> Proportion | Big Salmon <br> Sonar <br> Count | Eagle-GSI <br> generated <br> estimate | Difference <br> from sonar <br> count | \% <br> Difference |
| :---: | :---: | :---: | :---: | :---: | ---: |
| 2005 | 0.108 | 5618 | 7342 | 1724 | $30.7 \%$ |
| 2006 | 0.097 | 7308 | 6075 | -1233 | $-16.9 \%$ |
| 2007 | 0.106 | 4506 | 3700 | -806 | $-17.9 \%$ |
| 2008 | 0.093 | 1431 | 3151 | 1720 | $120.2 \%$ |
| 2009 | 0.169 | 9261 | 11032 | 1771 | $19.1 \%$ |
| 2010 | 0.117 | 3817 | 3745 | -72 | $-1.9 \%$ |
| 2011 | 0.092 | 5156 | 4672 | -484 | $-9.4 \%$ |
| 2012 | 0.067 | 5156 | 3402 | -1754 | $-34.0 \%$ |
| 2013 | 0.066 | 3242 | 1892 | -1350 | $-41.6 \%$ |
| 2014 | 0.024 | 6321 | 1520 | -4801 | $-76.0 \%$ |
| Mean | 0.094 | 5282 | 5390 | 108 | -0.03 |
| SD | 0.029 | 2319 | 2703 | 1437 | 0.53 |


[^0]:    ${ }^{1}$ Radio tagged Chinook were documented entering a tributary and subsequently retreating to the mainstem river and continuing their migration further up the system. Since the sonar station is located 1.5 km upstream from the confluence of the Yukon River the presence of straying Chinook could be expected.

[^1]:    ${ }^{2}$ Scale age analysis was conducted for DFO Whitehorse by the Pacific Biological Station, fish ageing lab in Nanaimo, British Columbia.
    ${ }^{3}$ European age format; e.g. 1.3 denotes a 5 year old fish with $1+$ years freshwater residence and 3 years marine residence.

[^2]:    ${ }^{4}$ Precision refers to the repeatability of a count between different individuals reading the same sonar file.
    ${ }^{5}$ Spawning escapement is the Eagle sonar count minus the catches in the U.S. above the sonar station and in the Canadian fisheries.

